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5 **Bird taxonomic and functional responses to land abandonment in** 6 **wood-pastures**

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50 **Abstract**

51 Wood-pastures are socio-ecological systems covering vast areas in Europe. Although used for graz-
52 ing and production of various forest goods, wood-pastures harbour a rich biodiversity and are usual-
53 ly considered as High Nature Value Farmlands. However, socio-economic pressures are driving the
54 transformation of these valuable landscapes from multi-functional, heterogeneous habitats to homo-
55 geneous areas through either intensification or land abandonment. We investigated how changes in
56 management intensity influence the taxonomic diversity, functional diversity and functional compo-
57 sition of birds in these landscapes using generalized linear models. In contrast to taxonomic diversi-
58 ty, functional diversity decreased significantly towards shrub-dominated and less heterogeneous ar-
59 eas related to the abandonment of grazing and/or understory management practices. Grassland and
60 generalist species, and associated guilds such as granivores, ground-nesters and ground-foragers are
61 almost absent less managed areas. On the other hand, shrub-dominated areas favour forest species,
62 particularly understory/canopy foragers and arboreal nesters, although the forest guild is still well-
63 represented in actively managed, heterogeneous areas. Our results indicate the abandonment of
64 wood-pasture management affects the prevalence of grassland and generalist species, leading to
65 functional diversity loss and potentially reduced ecosystem functioning. We suggest non-intensive,
66 active management is needed to maintain habitat heterogeneity and canopy openness, enhancing
67 trait diversity in wood-pastures.

68 **Keywords:** functional diversity; breeding birds; wood-pasture; management intensity;
69 montado/dehesa; socio-ecological systems

70

71 **1. Introduction**

72 Wood-pastures are social-ecological landscapes that have been shaped by various land-use
73 regimes prevailing in much of Europe (Hartel and Plieninger 2014). They usually have high
74 economic value provided by a multi-functional management that may include livestock grazing,
75 cork extraction, timber production, crop cultivation and tree pruning for firewood and charcoal
76 (Moreno et al. 2018). Many wood-pastures conciliate economic value with the maintenance of rich
77 biodiversity, and as such are considered High Nature Value Farmlands (Pinto-Correia and Ribeiro
78 2012). This richness is particularly evident in wood-pasture landscapes which are spatially
79 heterogeneous due to the availability of habitats such as riparian galleries, hedgerows, shrubby
80 patches and olive orchards (Leal et al. 2016; Oksuz et al. 2020). The maintenance of the multiple
81 values of wood-pastures depends on a balanced management of the landscape (Plieninger et al.
82 2015). However, recent social and economic changes in Europe are transforming traditional wood-
83 pasture management through intensification or abandonment (Bergmeier and Roellig 2014), putting
84 the balance between natural and economic values at risk.

85 Land-use intensification induces habitat homogenization and causes biodiversity loss in
86 wood-pastures (Flynn et al. 2009). Land abandonment also affects many wood-pastures (Godinho et
87 al. 2016) although its consequences are comparatively less studied than intensification (Estel et al.
88 2015). Land abandonment, which may have natural (Rey Benayas et al. 2007) and socio-economic
89 drivers (Levers et al. 2018), often leads to areas with little use beyond extensive grazing or even a
90 complete ceasing of all exploitation, resulting in less heterogeneous habitats with dense shrubby
91 vegetation (Rey Benayas et al. 2010). These landscape changes can influence biodiversity and
92 ecosystem processes, although the effects vary depending on the taxon and geographical region
93 (Queiroz et al. 2014). For instance, the loss of open areas can reduce habitat heterogeneity and,
94 consequently, the diversity of feeding and nesting resources for birds (Sirami et al. 2009). The
95 subsequent loss of bird species changes the trait composition of assemblages, which may in turn
96 affect ecosystem functioning and services (Hooper et al. 2005). Thus, assessing how trait

97 composition responds to land abandonment, particularly those traits that simultaneously influence
98 species responses and functions (i.e. “response and effect traits”; Luck et al. 2012; Díaz et al. 2013),
99 can reveal how biodiversity and ecosystem functions may be affected by this type of land use
100 change (Mouillot et al. 2013). This knowledge is crucial to inform managers and decision-makers
101 about potential negative consequences of such changes and to develop adequate strategies to
102 minimize them (Wood et al. 2015).

103 In this study, we aimed to explore how changes in habitat structure driven by different
104 management intensities influence bird taxonomic diversity, functional diversity and trait
105 assemblages in wood-pastures across the Iberian Peninsula and North Africa. Birds, being good
106 indicators of environmental change (Sekercioglu 2006), are likely to represent key ecosystem
107 services in wood-pastures, including seed dispersal, pest regulation or pollination (Pons and Pausas
108 2007; Ceia and Ramos 2016). We related taxonomic and functional indices to habitat structure,
109 characterized by a set of variables including understory cover and height, tree density and habitat
110 heterogeneity, as indicators of a management intensity gradient ranging from active management
111 towards land abandonment. Specifically, we aimed to answer the following questions: i) How do
112 changes in management intensity gradient influence the taxonomic and functional diversity of birds
113 in wood-pastures? ii) Which specific traits drive the responses of birds to the habitat along the
114 management intensity gradient?

115 **2. Methods**

116 **2.1 Study Area**

117 Our study encompassed much of Portugal, Spain and Morocco, in areas dominated by a
118 wood-pasture system distributed across the western Mediterranean, known as *montado* in Portugal,
119 *dehesa* in Spain and *azaghar* in some regions of North Africa (Fig. 1). The woody plant
120 composition of the sampled areas is dominated by cork oak (*Quercus suber*), sometimes co-

121 occurring with other oaks (e.g. Algerian oak *Q. canarensis*, holm oak *Q. rotundifolia*, Pyrenean oak
122 *Q. pyrenaica*), pine trees (e.g. stone pine *Pinus pinea*, maritime pine *P. pinaster*), and other tree and
123 shrub species (e.g. mastic tree *Pistacia lentiscus*, wild olive trees *Olea europaea* var. *sylvestris*,
124 strawberry tree *Arbutus unedo*, etc.). The most common management practices in the study area are
125 livestock grazing with various degrees of intensity, cork extraction, cropping and pruning
126 (Berrahmouni et al. 2007; Moreno et al. 2018). Moderate to intensive grazing is prevalent in
127 lowland wood-pastures, whereas shrub encroachment tends to occur in more rugged areas (Bugalho
128 et al. 2009; Bugalho et al. 2011). The annual average temperature ranges between 11-18°C and the
129 annual rainfall is 410-910 mm in the study area (<http://www.worldclim.org/>).

130 **Fig. 1** Map of the study areas located in Portugal, Spain and North Africa. Circles represent the
131 sampled wood-pastures in Portugal (N=17), Spain (N=13) and in Morocco (N=7).

132

133 **2.2 Bird sampling**

134 Bird sampling was performed during the spring of 2011 using five-minute bird point counts
135 (Bibby et al. 2005). In total, thirty-seven wood-pastures with a minimum of 50 hectares and at least
136 10 km apart, were sampled across Iberia (17 in Portugal, 13 in Spain) and North Africa (7 in
137 Morocco). We were able to investigate the biodiversity patterns of birds across Europe and North
138 Africa given the similarity of the bird species assemblages in both Europe and North Africa wood-
139 pastures (Correia et al. 2015a). Fifteen sampling stations were set up in each of these wood-pasture
140 areas, at least 200 m apart and 100 m or more from the edge. Each station was visited twice, once
141 during the early half (1 April to 15 May) and once during the late half (16 May to 20 June) of the
142 breeding season. The same observer performed all the counts, always during periods of peak bird
143 activity (controls in the morning and in the late afternoon for each area) and avoiding rainy and
144 windy conditions and areas recently harvested for cork (Godinho and Rabaça 2011). All birds

145 detected visually or acoustically were recorded and their distance to the observer was estimated.
146 Birds detected more than 100 m away from the observer and over-flying birds were excluded from
147 the analysis since they may not be relevant to the studied habitat. The total abundance of each
148 species in each area was defined as the maximum sum of individuals detected in the fifteen
149 counting stations for any of the two controls, as this represents the minimum number of birds
150 present in that area (Bibby et al. 2005). Bird abundance data is presented in Online Resource 1.

151

152 **2.3 Trait data**

153 We obtained data on six response and effect traits (Luck et al. 2012; Hevia et al. 2016) for
154 54 recorded species to analyse the functional diversity and functional composition of bird
155 assemblage. The traits considered for analysis include: habitat guild, feeding guild, foraging strata,
156 nest type, wing aspect ratio and body mass (Online Resource 2). These traits were selected based on
157 a priori hypotheses regarding their role in determining bird responses to habitat structure (Online
158 Resource 3).

159 **2.4 Environmental data**

160 Habitat structure was characterized with a set of variables including understory cover and
161 height, tree density and habitat heterogeneity (Online Resource 4). These habitat variables were
162 estimated in the same circle plots where bird census was performed. These variables were selected
163 as indicators of a management gradient ranging from areas with active human management (e.g.
164 regular grazing, shrub removal) to sparsely used areas (e.g. occasional grazing, no shrub removal)
165 resembling a process of land abandonment. Herb cover and shrub cover (% of ground cover), herb
166 height (in 5 cm classes up to 25 cm) and shrub height (in 25 cm classes up to 150 cm) were visually
167 estimated by the same observer. Tree density (number of trees per hectare) and canopy cover (%
168 ground cover covered by tree crowns) were estimated visually using aerial images available from
169 **Google Earth v7.1 for the year 2011**. A habitat heterogeneity index was obtained using six variables

170 - herb cover, herb height, shrub cover, shrub height, tree density and canopy cover - applying the
171 formula of “max.value–min.value/mean value” for each habitat variable. The resulting values were
172 summed to obtain a single heterogeneity value for each wood-pasture (Rotenberry and Wiens 1980).
173 Finally, we performed a Principal Components Analysis (PCA) combining variables of herb and
174 shrub cover, herb and shrub height, tree density and habitat heterogeneity to reduce the number of
175 variables and avoid collinearity (Dormann et al. 2013). The first and the second axis of the PCA
176 were used to represent the main habitat management gradients in the subsequent analysis (Table A1,
177 Online Resource 5). The first PCA axis mostly represents ground vegetation structure ranging from
178 actively managed herb dominated and more heterogeneous areas to shrub dominated and less
179 heterogeneous areas, where the management intensity is lower. The second PCA axis represents a
180 gradient from sparsely treed and heterogeneous areas to densely treed and less heterogeneous areas.
181 Shrub cover and height are lower in these densely treed areas due to more intensive management.
182 (Figure A1, Online Resource 5).

183 **2.5 Data Analysis**

184 We estimated bird taxonomic diversity using species richness and the Shannon diversity
185 index, and estimated functional diversity with multi-trait functional richness (FRic), functional
186 dispersion (FDis) and functional evenness (FEve) indices (Villéger et al. 2008). Before each index
187 was calculated, we assessed the correlation between traits using Spearman correlations to avoid
188 biases in calculation of functional diversity indices (Lepš et al. 2006). There were no correlations
189 between traits ($p>0.05$), so they were all included in calculations and given equal weight. We also
190 calculated the community weighted means (CWM) index of each trait to test how individual trait
191 composition responded to changes in habitat structure.

192 All analyses were performed in R, version 3.5.2 (R Core Team 2019). Species richness and
193 Shannon index were calculated using the “vegan” package (Oksanen et al. 2016), and functional
194 diversity and functional composition indices with package “FD” (Laliberté et al. 2014). Later, we

195 tested the relationship between biodiversity indices and the management intensity gradients
196 represented by PCA axes using generalized linear models. Adjusted R-squared values were
197 calculated for each model using “rsq” function in “rsq” package (Zhang 2018). Finally, we
198 performed a Detrended Correspondence Analysis (DCA) using “vegan” package to confirm that we
199 were able to test the biodiversity of birds across Europe and North Africa (Fig. A2 and Table A2,
200 Online Resource 5). All figures were produced using “ggplot2” (Wickham 2016).

201 **3. Results**

202 **3.1 Taxonomic diversity**

203 There were no evident changes in species richness or Shannon diversity in relation to the
204 first axis representing management intensity ($p>0.05$). Both species richness and diversity
205 decreased towards more densely treed and less heterogeneous areas (second axis representing
206 management intensity), but neither trend was significant ($p>0.05$) (Fig. A3 and Table A3, Online
207 Resource 5).

208

3.2 Functional diversity

209 Functional diversity showed a significant response to management, generally decreasing
210 towards lower management intensity. FRic showed a nearly-significant decrease in relation to PCA
211 axis1 ($p=0.08$), and thus towards areas with less management, where shrub cover and shrub height
212 are higher (Fig. 2a). Fdis ($p<0.001$; $R^2=-0.29$) and FEve ($p\leq 0.01$, $R^2=-0.21$) also decreased towards
213 shrub-dominated and less heterogeneous areas (Figs. 2b and 2c), and in both cases the decrease was
214 significant. No significant associations were detected between FD indices and the second PCA axis
215 ($p>0.05$). All diversity data is available in Table A4, Online Resource 5.

Fig. 2 Relation of bird functional diversity to habitat structure, represented by the first PCA axis.

All variables decrease towards shrub dominated, less managed areas, but the trend is significant

only in (b) and (c), represented by solid trend lines. See Table A4, Online Resource 5 for test statistics of linear models.

3.3 Functional composition

216 We observed significant variations in feeding guild, foraging strata, nest type and habitat use
217 traits of birds along the management gradient represented by the first PCA axis (Table A5, Online
218 Resource 5). CWM of grassland species ($p < 0.001$; $R^2 = -0.56$) and generalists ($p \leq 0.05$; $R^2 = -0.08$)
219 significantly decrease towards shrub dominated areas, while forest species ($p < 0.001$; $R^2 = 0.46$)
220 show the opposite trend. However, it should be noted that forest species are still well represented in
221 more open and heterogeneous areas, where they compose approximately half of the observed bird
222 communities, whereas grassland specialist species are often absent in shrub dominated areas (Fig.
223 3a).

224 CWM of granivores ($p \leq 0.05$; $R^2 = -0.11$) decreases significantly towards less managed,
225 shrub-dominated areas. However, the CWM of omnivore and insectivore species ($p > 0.05$) did not
226 vary significantly in relation to any of the habitat variables, indicating their relative abundance
227 remains constant independently of habitat structure (Fig. 3b). The relative abundance of ground-
228 foragers ($p < 0.001$; $R^2 = -0.39$) also decreases with higher shrub cover in areas tending towards
229 abandonment, while species foraging in the canopy ($p < 0.001$; $R^2 = 0.34$), midhigh ($p \leq 0.01$; $R^2 = 0.20$)
230 and understory ($p < 0.001$; $R^2 = 0.35$) benefit from the abundant woody vegetation available in these
231 areas (Fig. 3c). Closed-ground nesters ($p \leq 0.05$; $R^2 = -0.21$) are better represented in open and
232 heterogeneous areas maintained by active human management. On the other hand, closed-arboreal
233 ($p \leq 0.01$; $R^2 = 0.15$) and open-arboreal ($p \leq 0.05$; $R^2 = 0.12$) nesting birds are more prevalent in less
234 managed areas with higher shrub cover and tree density (Fig. 3d).

235 Only wing aspect ratio ($p \leq 0.05$; $R^2 = -0.08$) decreased with intensive shrub management, as
236 suggested by the significant negative relation to the second PCA axis (Fig. A4, Online Resource 5).

237 We did not observe significant variations in body mass, tree/cavity and bare ground nesting types in
238 relation to habitat variables (Table A5, Online Resource 5).

239 **Fig. 3** Community weighted means (CWM) of functional traits across the management gradient.
240 CWM of generalists and grassland species (a) granivores (b) ground-foragers (c) closed-ground
241 nesters (d) decreases towards shrub dominated, less managed areas, while forest species (a), canopy
242 and understory foragers (c) and arboreal nesters (d) benefit from shrublands. Solid and dashed lines
243 indicate significant ($p \leq 0.05$) and non-significant trends, respectively. See Table A5, Online
244 Resource 5 for test statistics of linear models.

245

246 **4. Discussion**

247 **4.1 Land abandonment leads to decreases in bird functional diversity, but not in taxonomic** 248 **diversity**

249 Our results show that changes in vegetation structure and habitat heterogeneity linked to
250 land abandonment influence trait-level diversity of birds during the breeding season, but do not
251 influence taxonomic diversity. Neither species richness nor Shannon diversity metrics responded to
252 changes in habitat structure, indicating the presence of diverse species assemblages across the
253 management gradient (Fig. A3, Table A3, Online Resource 5). However, we found a significant
254 decline of functional dispersion and evenness toward less managed areas (Fig. 2), suggesting that
255 trait assemblages are more similar and less evenly distributed in wood-pastures within areas
256 undergoing active management (Hillebrand et al. 2008; Crowder et al. 2010). Studies report
257 contrasting trends in species and trait-level responses to various management strategies, suggesting
258 the need to explore different dimensions of biodiversity to understand the complex relation of
259 species to the ecosystem (Devictor et al. 2010). Our results support this view and suggest more
260 pronounced trait-level responses to changes in wood-pasture structure than those observed at the

261 species-level, mostly due to the loss of traits of grassland birds (Fig. 3a). Variations in niche
262 structure due to different management strategies are likely to explain the observed patterns, which
263 can benefit or impair species with specific traits without necessarily leading to changes in
264 taxonomic diversity (Gagic et al. 2015).

265 Land abandonment in wood-pastures is often characterised by a reduction in grazing and the
266 absence of activities commonly used to improve grazing potential, namely shrub removal, resulting
267 in higher tree and shrub densities, denser canopies and a decrease in habitat heterogeneity (Castro
268 and Freitas 2009; Oldén et al. 2017). Grassland species usually depend on open mosaics of habitats
269 to feed and nest (Reino et al. 2010), therefore the presence of denser and taller woody vegetation is
270 likely to reduce the availability of suitable conditions for these species (Sirami et al. 2007; Spitzer
271 et al. 2008). Our results suggest this is indeed the case in wood-pastures as the relative abundance
272 of grassland species decreased towards areas with denser and taller shrubby vegetation (Fig. 3a).
273 Grassland bird populations have suffered severe declines in recent decades all across Europe mostly
274 because of land-use intensification (Butler et al. 2010), but our findings underline land
275 abandonment may be another factor negatively affecting this guild. Traits such as ground nesting,
276 ground feeding and granivory are often available only in open areas, also decreased towards
277 abandoned areas (Fig. 3). These changes are likely due to the presence of dense shrubby vegetation
278 in these areas, which can substantially restrict seed availability, predator detectability and the
279 availability of suitable ground foraging sites for these species (Vickery and Arlettaz 2012).

280 In contrast, our results show forest species may benefit from land abandonment (Fig. 3a) and
281 similar results observed in wood-pastures across Europe (Sirami et al. 2009; Nikolov et al. 2011;
282 Jakobsson et al. 2018). Dense shrublands are known to be suitable for forest species such as
283 *Erithacus rubecula*, *Sylvia atricapilla* and *Troglodytes troglodytes* by providing feeding resources
284 and reducing predation (Santana et al. 2012). Furthermore, the availability of nesting sites for
285 arboreal nesters is likely to increase in densely wooded areas, as our results present (Fig. 3). We

286 also observed significant differences in birds linked to different foraging strata (Fig. 3), which is
287 informative of the way birds use the habitat under different management strategies (Martin and
288 Possingham 2005). The abundance of understory and canopy foragers increases whereas ground
289 foragers decrease in less managed areas, suggesting that land abandonment leads to changes in
290 different foraging strata use.

291 There may be important consequences to wood-pasture ecosystem dynamics related to
292 vertical shift of birds towards higher foraging strata in less managed areas. For example, many oak
293 pest species spend at least part of their life-cycle on or under the ground (Ceia and Ramos 2016).
294 Ground foraging birds may thus play an important role controlling these pests, but this role may be
295 limited in areas with dense ground vegetation, as our results further suggest.

296 We also observed that bird assemblages in more homogeneous areas with higher tree density
297 and less shrubs tend to feature more birds with a lower wing aspect ratio (Fig. A4, Online Resource
298 5). Lower wing aspect ratio represent shorter and more rounded wings allowing for better
299 manoeuvrability in densely treed habitats (Vanhooydonck et al. 2009) but are less suited for longer
300 distance flights.

301 **4.2 Active management maintaining habitat heterogeneity supports functionally diverse bird** 302 **communities in wood-pastures**

303 Our results provide evidence for the important role of management in maintaining bird
304 functional diversity in wood-pastures. Extensive grazing acts to maintain the characteristic habitat
305 heterogeneity of wood-pastures which is crucial to provide diverse feeding and nesting resources
306 for birds (Tews et al. 2004; Leal et al. 2019) particularly during the breeding season (Mag and Ódor
307 2015). Grassland species are especially dependent on resources that are only available in more open
308 and heterogeneous areas, which are usually sustained by extensive grazing (Reino et al. 2010).
309 Generalists also seem to increase with active management, whereas forest species dominate the

310 assemblage in the denser habitats that results from land abandonment. However, we underline that
311 forest species were also present in open managed areas, where they comprised approximately half
312 of the bird community, whereas grassland species are often absent in densely vegetated areas (Fig.
313 3a). This result indicates that by maintaining habitat heterogeneity, actively managed areas can
314 provide a wide range of niches, inclusively for most forest species.

315 While active management seems necessary to maintain heterogeneous wood-pasture
316 landscapes, the intensity of management also plays a key role. There is evidence that in highly
317 exploited systems, reducing management intensity level can be advantageous providing more
318 suitable and connected habitats for species narrower niche requirements (Queiroz et al. 2014). On
319 the other hand, studies focusing on multiple species responses to land abandonment also report the
320 potential threats of this change in land-use types to the overall biodiversity value of managed
321 habitats (Horák et al. 2018a). Specifically, land abandonment may ultimately lead to the functional
322 homogenization of biological communities in Mediterranean habitats (Clavero and Brotons 2010)
323 due to a loss of habitat heterogeneity, as our results also indicate.

324 It is obvious that both habitat intensification and abandonment can have adverse
325 consequences for the biodiversity of wood-pastures, whose preservation depends on the
326 maintenance of habitat heterogeneity emerging from multi-purpose management strategies
327 (Mönkkönen et al. 2014; Roellig et al. 2016). As more traditional management strategies struggle to
328 maintain the economic sustainability of wood-pastures, the challenge for the future is to find simple
329 and inexpensive management strategies that conciliate economic and natural values in wood-
330 pastures.

331 Ultimately, our results propose that maintaining active management may lead to higher
332 response trait diversity of birds in wood-pastures. Lastly, we should underline that the traits we
333 analysed are both “response and effect traits”, implying that many of the trait-level changes

334 observed may also reflect greater effect on the ecosystem services (de Bello et al. 2010; Torralba et
335 al. 2016).

336 **5. Conclusions**

337 We show substantial shifts in bird assemblage functional diversity and composition, in
338 response to vegetation structure and habitat heterogeneity changes associated with land
339 abandonment in wood-pastures. These changes seem to be unfavourable for grassland and
340 generalist species with life-history traits such as ground nesting, granivory and ground feeding,
341 ultimately leading to loss of bird functional diversity and potentially hampering the range of
342 ecosystem services that birds can provide. In many regions, wood-pastures are undergoing a
343 transformation from multi-functional heterogeneous systems to more homogeneous habitats, due to
344 a reduction of management resulting from multiple socio-economic pressures (Hartel et al. 2015).
345 The abandonment of traditional management, and the subsequent encroachment of woody
346 vegetation, are increasingly threatening the biodiversity of many European habitats (Queiroz et al.
347 2014). Balancing the natural and economic values of human-modified habitats is one of the main
348 challenges in improving conservation efforts (Graves et al. 2007; Landis 2017) and our results
349 suggest that maintaining non-intensive, active management may be a sustainable answer to this
350 challenge in wood-pastures.

351

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357 **Conflicts of interest**

358 The authors have no relevant financial or non-financial interests to disclose.

359 **Ethics approval**

360 Not applicable.

361 **Consent to participate**

362 Not applicable.

363 **Consent for publication**

364 Not applicable.

365 **Availability of data and material**

366 All data generated or analysed during this study is included in Supplementary Information files
367 (Online Resources 1 and 2).

368 **Code availability**

369 Not applicable.

370 **Authors' contributions**

371 **Duygu P. Oksuz:** Conceptualization, Data Curation, Formal analysis, Investigation, Methodology,
372 Writing-original draft, Writing-review&editing.

373 **Jorge M. Palmeirim:** Conceptualization, Methodology, Supervision, Funding
374 acquisition, Writing-review&editing.

375 **Ricardo A. Correia:** Conceptualization, Data Curation, Investigation, Methodology, Funding
376 acquisition, Supervision, Writing-review&editing.

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380

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